ASSESSMENT OF HEAVY METALS DISTRIBUTION IN THE SURFACE SEDIMENT OF KEMAMAN RIVER, TERENGGANU

VISHALINI B.MARAN¹, DORINDA ANTHONY ANTHONY DASS¹, LAVANNIA RAVIKUMAR¹, NUR SYAMIMI IZYAN ZAINI¹, NUR ALIAH SYAKIRAH ROSLI¹, WAN NUR IZWANI MIOR BAHARUDIN¹, NAJAH KARIMAH MUSTAFFA¹ AND ONG MENG CHUAN^{1,2,3*}

¹Faculty of Science and Marine Environment, Universiti Malaysia Terengganu, 21030 Kuala Nerus, Terengganu ²Institute of Oceanography and Environment, Universiti Malaysia Terengganu, 21030 Kuala Nerus, Terengganu ³Ocean Pollution and Ecotoxicology (OPEC) Research Group, Universiti Malaysia Terengganu, 21030 Kuala Nerus, Terengganu

*Corresponding author: ong@umt.edu.my

Abstract: In the present day, rivers are used by people as water sources for their daily agricultural, industrial and cleaning purposes. However, industrial development and human activities are rapidly impacting the river's catchment area, which can cause contamination of heavy metals in the water and sediment. The Kemaman River is one of the rivers that may be affected by these pollutants. Thus, the concentration of selected heavy metals such as copper (Cu), zinc (Zn), cadmium (Cd), lead (Pb), arsenic (As) and mercury (Hg), plus the pollution status was determined in the surface sediments of the Kemaman River. A total of 46 sediment samples were collected along the river. Teflon Bomb digestion method with 2.0mL of mixed acid (concentrated nitric acid, hydrochloric acid and hydrofluoric acid with ratio 3:3:1) was performed followed by analyzation using Inductively Coupled Plasma Mass Spectrometry (ICPMS). The status of heavy metals contamination in the study area was determined by the index of geoaccumulation (I_{eco}) and pollution load index (PLI). The calculated I value suggested that the sediments in the Kemaman River were categorized under unpolluted to polluted category. However, the calculated PLI suggested that the study area is polluted with heavy metals. The correlation of the sediment mean grain size and the selected heavy metals was strongly correlated. The concentration of metals increased with the rising mean grain size, which suggests their association with the fine fraction of the sediment.

Keywords: heavy metals, sediment, Kemaman river, pollution, ICPMS

Introduction

The Kemaman River discharges freshwater of about 80 m³/s during the southwest monsoon season and 500 m³/s during the northeast monsoon season (Kamaruzzaman et al., 2002; Yusof, 2018). Industrial growth and human activities are increasingly impacting the river catchment area. A research carried out by Kamaruzzaman et al. (2002) found that the seawater intrusion into the Kemaman River is limited to a distance of 10 km upstream from the river mouth. Based on the collected data for 2020 obtained from the Malaysian Meteorological Department, the northeast monsoon season with strong winds and long frequency periods with a mean rainfall ranged from 141 to 614 mm, was from November to January. On the other hand,

the southwest monsoon season with low rainfall occurred from April to June (Kamaruzzaman *et al.*, 2002).

Heavy metal is an element or metalloid with high atomic weight and its density is five times more than water. Heavy metals are hazardous due to their bioaccumulation and toxicity behaviour (Tchounwou *et al.*, 2012). Another unique characteristic of these elements is that heavy metals cannot be degraded or destroyed (Ahmadipour *et al.*, 2014). In the environment, heavy metal element concentration with less than 10 ppm, is considered as trace element due to its presence in trace concentration (Kabata-Pendias, 2010). Heavy metals may settle down and accumulate in the marine sediment (Mansour *et al.*, 2013; Ong *et al.*, 2015).

Pollution which occurs in the marine environment due to natural and anthropogenic activities by heavy metals is a worldwide problem (Zhu et al., 2013; Tang et al., 2014). Generally, the marine environment is suffering from human activities such as industrialization, agricultural and other activities. Due to anthropogenic activities such as fishing and industrial estate activities, heavy metal pollution might have affected the Kemaman River. However, some metals are essential in ecological function (Suresh et al., 2012), mainly to aquatic organisms. Since they are not removed from the water, metals can circulate in water bodies and settle down in the sediment bed or enter the food chain (Rai et al., 2015). Changes in metal disposal into the surrounding environment can affect the organisms and plants if the metals are higher than the ambient level due to their toxicity level.

This study aims to determine the distribution of heavy metal in surface sediment of the Kemaman and Chukai Rivers. Hence, the current pollution status on heavy metals in surface sediment of the Kemaman and Chukai Rivers is defined. The current finding may also provide a reference for future research.

Materials and Methods

Kemaman district is located at the latitude of 4°13'60" N and longitude of 103°25'59" E, approximately 180 km southward of the Kuala Terengganu district. The Kemaman coastal area is a flat lowland and activities focus dominantly on fishing activities, local economic activities and farming. Two main rivers flow within this region, namely the Kemaman and Chukai Rivers. These two rivers are separated at the upstream and joined at the downstream, sharing the same estuary, known as the Kemaman River estuary. The Kemaman River has a greater discharge, mean annual discharge and is larger than the Chukai River.

Throughout this study, 33 sampling points were along the Kemaman River and 13 sampling points were along the Chukai River, with 500m distance between each point. The surface sediment was collected using Ponar grab. The sampling activities were accomplished in October 2019. Figure 1 shows the sampling stations along the Kemaman and Chukai Rivers



Figure 1: Map of sampling points in the Kemaman and Chukai Rivers

The collected sediments were digested using the method by Kamaruzzaman *et al.* (2010) and Ong *et al.* (2016) to obtain the heavy metal content. A total of 0.05 g of sieved sediment was weighed using an analytical

balance and transferred to a Teflon beaker. Next, 2.0 mL of mixed concentrated acids, which are nitric acid (HNO_3), hydrochloric acid (HCl) and hydrofluoric acid (HF) with a ratio of 3:3:1, respectively, were added to the sediment and

heated in the oven for 8 hours at 100 °C. The digested samples were then cooled to room temperature, transferred into centrifuge tubes and milli-Q water was added up to 10 mL. Samples were kept in the freezer until further analysis. The concentrations of heavy metals were determined using Inductively Coupled Plasma Mass Spectrometry (ICP-MS) (Ong *et al.*, 2015) and recorded in the Microsoft Excel spreadsheet. The accuracy of the digestion method was checked against blank samples and Standard Reference Material, SRM 1646a Estuarine Sediment.

For the particle size analysis, two methods were used since the samples comprise of sandy and fine particle samples. For fine particle samples, a total of 2g sample was diluted with distilled water and calgon solution was added in the samples to disperse bonded particles in the sediment by organic matter. After 24 hours, the sample was measured using the Particle Size Analyzer laser diffraction method (Joseph et al., 2020). For sandy particles, the samples were sieved using sieve shaker with 13 series sieving from the top to bottom of 4000 μ m, 2800 μ m, 2000 μ m, 1400 μ m, 1000 μ m, 710 μ m, 500 μ m,

355 μ m, 180 μ m, 125 μ m, 90 μ m, 63 μ m and less than 63 μ m. The sieve shaker was shaken for 15 minutes. The weight of the samples in each sieve was recorded (Joseph *et al.*, 2020). Next, GRADISTAT software was used to calculate the mean size particle and determine the type of sediment classification.

Results and Discussion

Generally, the sediment mean size value gives a simple indication of the degree of the force by water or wind which transports the sediment grain (Durán et al., 2011). Grain size is vital to identify the sediment particles entrainment, transportation and deposition (Singovszka et al., 2016). In this study, surface sediments vary from very coarse sand to coarse silt at both the Kemaman and Chukai Rivers. Increasing in mean size value indicates a decrease in grain size and vice versa. Figure 2 shows the sediment mean size value at each sampling station. The higher mean size value in the study area was observed at KMM-33 (-0.47Æ) which indicates very coarse sand due to the high energy in the environment (Sallenger and Richmond, 2011). The lowest mean size was found at KMM-28 (5.87Æ) with indicates coarse silt.



Figure 2: Mean grain size across all stations at the Kemaman and Chukai Rivers

Stronger water current in a particular study area makes the sediment coarse. This is due to the high velocity of the river water flows which transports the fine sand, and leaves the coarser sand on the river bed. Additionally, a high amount of suspended sediment is transported to the estuary and settled down during tidal (Kamaruzzaman *et al.*, 2002).

Table 1 shows the average concentration of heavy metals in the Kemaman and Chukai Rivers sediment and Figure 3 shows the distribution of metals in the sediment of the Kemaman and Chukai Rivers. The concentration of Cu shows high level at the downstream zone compared to upstream zone (Figure 3). The mean value recorded for the concentration of Cu was $15.1\pm11.8 \ \mu g/g dry weight$. The lowest concentration of Cu was recorded at KMM-11 with 2.40 μ g/g dry weight whereas high concentration of Cu was found at KMM-03, with 45.6 μ g/g dry weight. Meanwhile, the average value for Zn was $60.2\pm59.4 \ \mu g/g dry weight$. The concentration of Zn was in the range of 1.77 to 210 μ g/g dry weight. Similar to Cu, the concentration of Zn was high at the downstream zone compared to upstream zone. The lowest concentration of Zn was recorded at the CK-09 and highest concentration at KMM-13.

The concentration of Cd was also high at the downstream zone compared to the upstream zone. In this study, the mean value for the concentration of Cd was $0.10\pm0.07 \ \mu g/g$ dry weight. The lowest concentration of Cd was recorded at the KMM-26 with 0.01 $\mu g/g$ dry weight, while KMM- 08 recorded the highest concentration of Cd, with 0.27 $\mu g/g$ dry weight.

The concentration of Pb varied between the sampling stations. The highest and lowest concentrations of Pb were 2.71 μ g/g dry weight at CK-13 and 65.8 μ g/g dry weight at KMM-08, respectively. Figure 3 shows the concentration of Pb in the surface sediment samples of the Kemaman and Chukai River. The concentration of Pb was high at downstream compared to upstream. In this study, the average value for concentration of Pb was 24.4±19.5 μ g/g dry weight.

The mean value for the concentration of As in the Kemaman and Chukai River sediment was $3.78\pm3.15 \ \mu g/g \ dry \ weight$. The highest concentration of As was recorded at KMM-03 and the lowest at CK-13. The values ranged from 0.64 to 14.2 μ g/g dry weight. The concentration of As showed a high concentration at the downstream zone compared to the upstream zone. On the other hand, the concentration of total mercury (T-Hg) was high at the downstream zone compared to the upstream zone. The mean value for the concentration of T-Hg was $0.01\pm0.05 \ \mu g/g$ dry weight. The lowest concentration of T-Hg was recorded at the KMM-32 with 0.01 μ g/g dry weight whereas KMM-13 recorded the highest concentration of T-Hg with 0.15 μ g/g dry weight.

	Average Concentration (µg/kg dry weight)						
Zones	Cu	Zn	Cd	Pb	As	T-Hg	
Downstream	19.6	87.4	0.14	35.6	6.04	0.08	
Midstream	18.3	71.1	0.10	26.1	3.84	0.06	
Upstream	10.1	36.3	0.07	16.4	2.39	0.03	

 Table 1: The average concentration of heavy metals in the surface sediment samples of the Kemaman and Chukai Rivers, Terengganu



Figure 3: The distribution of metal concentration in the surface sediments of the Kemaman and Chukai Rivers, Terengganu using ArcGIS software

The grain size distribution of the sediments play a vital role in controlling the concentration and distribution of studied heavy metals (Sany *et al.*, 2013). This is due to the mobility of heavy metals in the sediments and water was controlled by several factors such as erosion, sedimentation, sediment type, water dynamics, urbanization, industrialization, river discharge and geochemical reactions (Liu *et al.*, 2011; Sany *et al.*, 2013). Throughout this study, the heavy metal concentration in the Kemaman and Chukai Rivers increased with a decrease in sediment size. Figure 4 shows the correlation value between heavy metals and sediment size. Lead showed a high relationship with sediment mean size which is r=0.76, followed by Zn (r=0.74) and T-Hg (r=0.72) which also indicates strong correlation.



Figure 4: Correlation between sediment size and heavy metal in the Kemaman and Chukai Rivers

The positive correlation coefficient between sediment size and heavy metals shows that the heavy metals are directly proportional to the particle size (Ong *et al.*, 2015). Coarse sediment size has a smaller ratio of surface area whereas fine sediment size has a larger surface area (Amin *et al.*, 2008; Ong *et al.*, 2013). Coarser sediments might contain a high composition of quartz and feldspar which lead to a lesser accumulation of heavy metals (Liu *et al.*, 2011). The fine sediment size gives a larger area of a binding site for heavy metals ion from the bottom layer of the water column (Yuan *et al.*, 2011; Bainbridge *et al.*, 2012).

Besides that, the correlation matrix explains that all the heavy metals have a good degree of relationship between each other. The positive correlations coefficient value was illustrated in Table 2. The strong correlation between

Pb and Zn (r=0.97) might be due to the city sewage effluents that accumulate Zn and Pb in sediment (Chakarvorty et al., 2015). The strong relationship between As and Cu (r=0.90) demonstrates a significant degree of positive correlation and might be influenced by similar possible sources (Singh et al., 2017). The significant relationship between T-Hg and Cd (r=0.70), As and Cd (r=0.77) were observed for their existence in the sediment samples and might be the characteristics of the natural origin for the heavy metals in the river sediment (Abdullah et al, 2015; Singh et al., 2017). Therefore, it is suggested that these metals may come from the same source, mutual dependence and have similar behavior in transportation which is similar to those found by Suresh *et al.* (2012), Bastami et al. (2014), Dou et al. (2013) and Singh et al. (2017).

Table 2: The correlation coefficient between heavy metals in the Kemaman and Chukai Rivers surface sediment

	Cu	Zn	Cd	Pb	As
Zn	0.88				
Cd	0.78	0.82			
Pb	0.86	0.97	0.88		
As	0.90	0.86	0.77	0.88	
T-Hg	0.79	0.86	0.70	0.86	0.84

Index of geoaccumulation (I_{geo}) was initially proposed by Muller (1969) to determine and define the metal contamination in sediment. Igeo is used to assess the contamination level of selected heavy metals in sediments by comparing with the background value or baseline (Muller, 1979; Mei *et al.*, 2011); plus, the results reflect the effect of lithogenic sources (Nobi *et al.* 2010; Sany *et al.*, 2013). The following equation is commonly used to calculate the metal contamination:

Igeo =
$$\log_2 \left(\frac{Cn}{1.5Bn}\right)$$

Where Cn is the measured concentration of elements "n" in the sediment. Bn is the background value for the metal "n" (Wedepohl, 1995). Factor 1.5 indicates the possible variations of the background values due to lithogenic effects (Muller, 1969). Table 3 shows the classification of Igeo.

Igeo class	Igeo	Description of sediment quality	
0	<0	Unpolluted	
1	0-1	Unpolluted to moderately contaminated	
2	1-2	Moderately polluted	
3	2-3	Moderately to strongly polluted	
4	3-4	Strongly polluted	
5	4-5	Strongly to extremely polluted	

Table 3: Classification of I_{geo} (Muller, 1969)

Based on the Igeo value, the trend of heavy metal pollution in the surface sediment of the Kemaman and Chukai Rivers are in the decreasing as follow, As > Pb > Cu > Cd > Zn

> T-Hg (Table 4). Cd and T-Hg are classified as unpolluted to moderately contaminated, while Cu, Pb and Zn are categorized as unpolluted to moderately polluted. Only As is found to be unpolluted to strongly polluted.

Figure 4: Pollution level of heavy metals based on Igeo classification (Muller, 1969) in the Kemaman and Chukai Rivers sediment

Metals	Mean	Min-Max	Category
Cd	-1.05	-3.73 - 0.83	
T-Hg	-1.43	-3.87 - 0.83	Unpolluted to moderately contaminated
Cu	-0.93	-3.16 - 1.09	
Pb	-0.59	-3.23 - 1.36	Unpolluted to moderately polluted
Zn	-1.28	-5.46 - 1.43	
As	-0.14	-2.23 - 2.25	Unpolluted to strongly polluted

The Pollution Load Index (PLI) gives a better understanding on the level of heavy metal contamination in the surface sediments (Tomlinson et al., 1980; Barakat et al., 2012). The PLI value is calculated using the following formula:

$$PLI=(CF_1 \times CF_2 \times CF_3 \times \dots \times CFn)^{1/n}$$

Where CF is the contamination factor for each metal. The contamination factor (CF) is the concentration of heavy metal in sediment which is divided by the background values (Rabee et al., 2011). "n" is the number of metal of the

studied metals. The PLI can be classified as polluted if the value is larger than 1 whereas PLI value less than 1 shows that no pollution occurs in the study area.

Figure 5 illustrates the PLI value of the studied heavy metals for each station. The range of PLI value was 0.20-3.42 and the average value was 1.14. This indicates that the Kemaman and Chukai River sediment are slightly polluted by the studied metals. The PLI values with less than 1 were mostly found in the upstream zone. Besides that, the PLI values with more than 1 were found at the downstream and midstream areas.

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Figure 5: PLI value for studied heavy metals for each station at the Kemaman and Chukai Rivers

Based on PLI analysis, the downstream and midstream zones tend to have high occurrence of contamination compared to the upstream zone. This is due to the high intense of human activities found at the downstream and midstream zones (Ong et al., 2012). Station KMM-03 showed the highest value of PLI due to the highest concentration in Cu and As among the sampling stations. Additionally, this station was also located within high fishing activity area. This is due to some sources of heavy metals which may have resulted by from the boat activities. Water can also transport high concentration of suspended sediments which contain metals due to the strong water current to the estuary areas (Kamaruzzaman et al., 2002).

Low contamination of metals occurred in the upstream zone. Higher hydrodynamic energy provides lesser resident time for metals to be absorbed in surface sediments (Sany et al., 2013). Thus, high portion of coarser grain size sediments was determined (Zhou et al., 2014). Metal ions are not easily absorbed by the water molecules, therefore, least of dissolved metals will be transported to the downstream zone (Sany et al., 2013). Thus, highest heavy metal contamination was obtained in surface sediments at downstream.

Conclusion

The sedimentological characteristics for the Kemaman and Chukai Rivers were categorized

as fine sand according to the mean size classification. The mean concentration of heavy metals was recorded as Zn (60.2±59.4 µg/g dry weight) > Pb (24.4 \pm 19.5 µg/g dry weight) > Cu (15.1±11.84 µg/g dry weight) > As $(3.78\pm3.15 \ \mu g/g \ dry \ weight) > Cd \ (0.10\pm0.07)$ μ g/g dry weight) > T-Hg (0.01 \pm 0.047 μ g/g dry weight). Generally, the downstream zone of the Kemaman and Chukai Rivers tends to have high concentration of heavy metals. Additionally, there is high relationship among each other in surface sediment from the Kemaman and Chukai Rivers. In conclusion, all metals were derived from the same source, plus all heavy metals are positively correlated with the sediment particles. The mean grain size might be important in controlling factors affecting the concentration and transportation of heavy metals. Coarse sediments has low ratio of surface area for metal ion absorption compared to fine sediments. According to average Igeo classification for individual metals, the contamination levels of heavy metals are not polluted. Some sampling stations were polluted with the studied metals. In contrast, based on the PLI value calculated, surface sediment of the Kemaman and Chukai Rivers can be classified as slightly polluted by metals. Therefore, continuous monitoring and further studies are recommended at the study area to determine the long term effects of metals on the riverine environment. This is to ensure no elevated pollution occurs which could affect the riverine ecosystem and the human pollution depending on the rivers.

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